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Spanish *Satureja montana* L. hydrolate: ecotoxicological study in soil and water non-target organisms --Manuscript Draft--

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Spanish *Satureja montana* L. hydrolate: ecotoxicological study in soil and water non-target organisms

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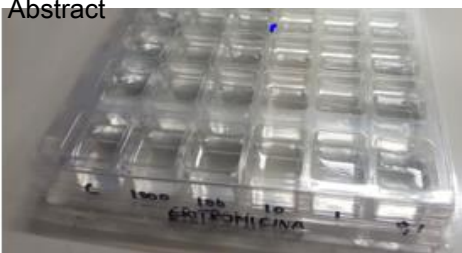
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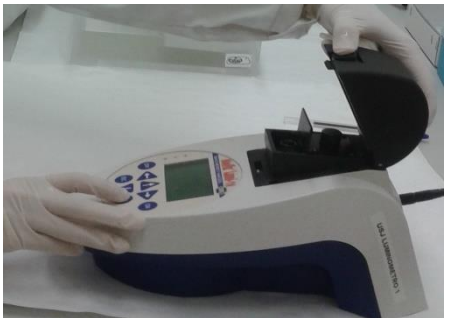
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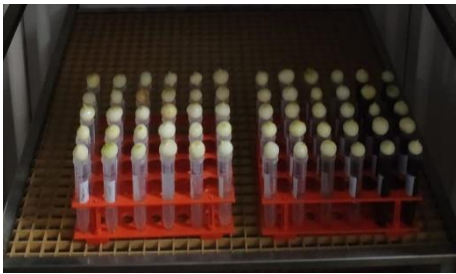
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Daphnia magna



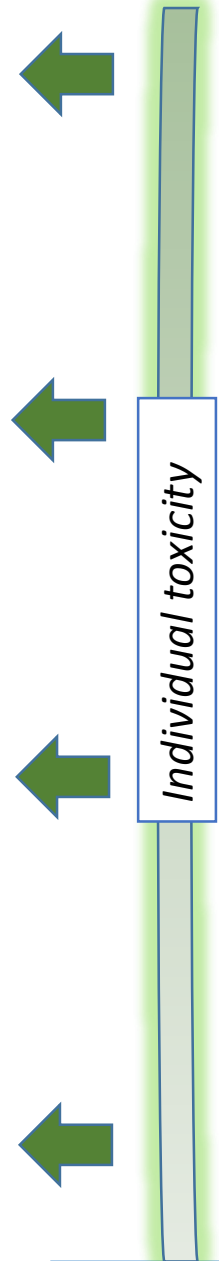
Vibrio fisheri



Allium cepa



Eisenia fetida



Individual toxicity

Acute toxicity



Satureja montana
hydrolate



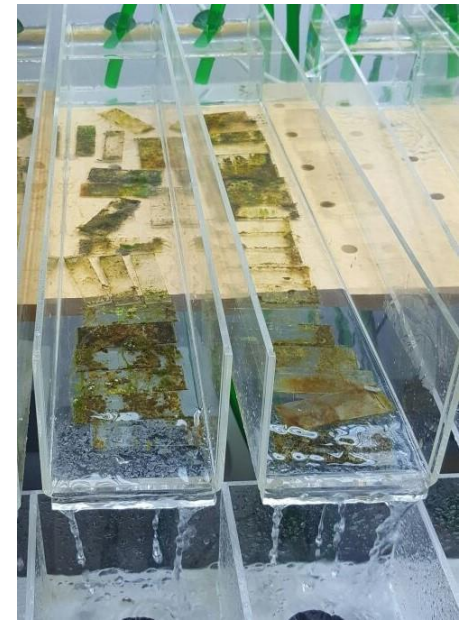
Community effects

Photosynthesis yield

Periphyton
communities



River microbial
community



HIGHLIGHTS

- *Satureja montana* hydrolate is highly phytotoxic on *Allium cepa* (LC₅₀ =0.05%)
- The hydrolate has a high ecotoxicity on *Daphnia magna* and *Vibrio fischeri* (LC₅₀<1%)
- *Eisenia fetida* is the single organism more resistant to the hydrolate (LC₅₀=4.25%)
- Communities of fluvial periphyton are also sensitive to hydrolate (LC₅₀=4.23%)
- Carvacrol (89%) and thymol (7%) are the main volatile compounds of the hydrolate

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Abstract

Despite the current popularity of herbal and flower hydrolates and the multiple applications found in the food, wood, pharmaceutical, pesticide, perfume, cosmetic, and aromatherapy industries, the effect of hydrolates in the environment is poorly known. This study evaluates the ecotoxicity of the *Satureja montana* L. hydrolate on water and soil bioindicators and also on the fluvial periphyton mesocosms for a more ecological point of view. The acute toxicity of the fresh water invertebrate *Daphnia magna*, the bacteria *Vibrio fischeri*, and the earthworm *Eisenia fetida*, was quantified as well as the phytotoxic effect on the plant *Allium cepa* L. Communities of river periphyton were used to study the impact of the hydrolate on the freshwater ecosystems. The taxonomic study of these communities revealed a rich diversity of diatoms. The hydrolate of *S. montana* showed a high ecotoxicity at very low percentages of hydrolate in all organisms tested and in the periphyton communities. The LC₅₀ varies from 0.05% to 4.25%, with a clear dose-dependent relationship. The effect of the hydrolate in decreasing order was: *A. cepa* > *D. magna* > *V. fischeri* > periphyton > *E. fetida*. The strong phytotoxic effect on *A. cepa* allows exploring possible uses of the hydrolate as a bio-herbicide. The Gas Chromatography Mass Spectrometry (GC-MS) characterisation of the hydrolate reveals 1 alcohol and 6 terpenes. Carvacrol (89.03%) and thymol (6.66%) are the volatile compounds found in the highest proportion, products that have multiple biological properties and known synergistic effects, which could explain the high bioactivity of the hydrolate. Our study suggests that *S. montana* hydrolate could impact different trophic levels of the river ecosystems and can affect the soil functions affecting earthworms due its powerful bioactivity on a wide range of non-target organisms even in complex communities such as the fluvial periphyton. Therefore, although hydrolates can become a good alternative to synthetic products, the use of these products is not free of environmental risks and their release to the environment should be evaluated.

29 **1. Introduction**

30 The *Satureja* L genus. (Lamiaceae Family) has numerous species widely distributed throughout the
31 world, most of them are shrubs or aromatic herbs with interesting biological properties (Saeidnia et
32 al., 2016). *Satureja montana* L., also called winter savory, is a perennial shrub with white or slightly
33 pink flowers (Lawless, 2002), industrially grown and appreciated for its medicinal, bactericidal and
34 fungicidal properties (Lopez-Cobo et al., 2015).

35 *Satureja montana* essential oil (EO) has a wide variety of pharmacological properties as
36 antimicrobial, antifungal, antioxidant, antispasmodic, antiviral, and antidiarrheal (Caprioli et al., 2019;
37 Jafari et al., 2018; Tepe and Cilkiz, 2016). This EO also has insecticidal (Tepe, 2015) and
38 nematocidal (Faria et al., 2016) properties. The EO from the same Spanish plant, *S. montana*,
39 recently have shown nematocidal properties as well (Navarro-Rocha et al., 2020).

40 During distillation of any plant biomass to obtain these EOs, a certain proportion of them becomes
41 dissolved in distillation or condensate water leading to a mixture named hydrolate, containing a
42 variable quantity of EOs and volatile, water-soluble, secondary active metabolites (Labadie et al.,
43 2015; Zheljzakov and Astatkie, 2011). Hydrolates are products of steam distillation and have been
44 considered a simple waste. This consideration has recently changed and many researchers have
45 found interesting biological properties of these aqueous fractions such as antifungal, antibacterial,
46 and antioxidant activities (Tornuk et al., 2011; Prusinowska et al., 2016; Maia et al., 2013; Franzener
47 et al., 2007; Di Vito et al., 2021).

48 Consumers' increasingly demand products with good environmental qualities, low toxicity and safe
49 for health so hydrolates appear as a good option to traditional treatments. Additionally, hydrolates
50 are readily available, inexpensive, and easy to produce (Tornuk et al., 2011). Hydrolates can act as
51 repellants, herbicides (Politi et al., 2020), and nematicides (Andres et al., 2018).

52 Hydrolates are widely used in the cosmetic industry and used as diluted in refreshing drinks (D'Amato
53 et al., 2018). The current research now also points to the use of hydrolates in the food industry to
54 prevent the growth of pathogenic microorganisms. Hydrolates also prevent organoleptic properties
55 of treated products and also possible risks to the health of consumers (Sagdic et al., 2013) such as
56 the appearance of carcinogenic by-products in chlorine-based sanitisers when reacting with organic
57 matter (Gil et al., 2009).

58 Although not well studied, the hydrolate of *Satureja* genus can act as a fungicide to prevent food
59 spoilage (Boyras and Ozcan, 2006) and have been proposed as an environmentally safer alternative
60 to prevent saprophytic and pathogenic fungi from damaging agri-food products (Boyras and Ozcan,
61 2006). Antimicrobial activity has also been reported for hydrolates of *Satureja* of different species
62 (Sagdic and Ozcan, 2003; Chorianopoulos et al., 2008; Giaouris et al., 2008; Sagdic et al., 2013;
63 Sahan and Tornuk, 2016). On the other hand, natural antioxidants obtained from hydrolates could
64 be a good alternative to synthetic antioxidants employed in the pharmaceutical and food industry,
65 possibly with little effect on organoleptic properties as described in the literature for hydrolates of
66 *Satureja* genus (Cabana et al., 2013).

67 Recently an Italian *S. montana* hydrolate showed important antimicrobial and antifungal properties
68 in a wide variety of microorganisms (Di Vito et al., 2021). Interestingly, the compounds of the
69 hydrolate were better able to inhibit microbial growth than EO from the same plant because they
70 showed activity at lower concentrations.

71 These findings highlight the potential of hydrolates as promising consumer products of commerce in
72 the food, wood, pharmaceutical, pesticide, perfume, cosmetics, and aromatherapy industries. For
73 example, the value of the hydrophilic EO fractions of aromatic oils that escape into the hydrolate was
74 estimated to be worth US\$50–100 million in India (Pangarkar, 2008).

75 Moreover, the exploitation of these by-products is an opportunity to extend their life cycle in
76 accordance with the circular economy guidelines and reduce the environmental impacts of waste
77 generation.

78 While potential uses are being studied, very little is known about ecotoxicity in non-target organisms
79 once these products reach the environment. Hydrolates could spread in the environment either as a
80 residue from EO extraction or as waste after consumption.

81 The first studies carried out on the ecotoxicity of hydrolates have shown that they are not at all
82 harmless for the environment. Studies on the ecotoxicity of hydrolates from *Lavandula luisieri*
83 (Rozeira) and *Artemisia absinthium* L. on the aquatic (Pino-Otin et al., 2019a) and edaphic
84 environments (Pino-Otin et al., 2019c; Pino-Otin et al., 2019b) show affection on individual
85 organisms as well as entire communities. Some studies also point to effects on fish: *Cinnamomum*
86 *zeylanicum* J. Presl hydrolate is toxic on carps (*Cyprinus carpio*) in aquariums (Gulec et al., 2013),

87 and *Lippia alba* (Mill.) N.E. Brown hydrolate showed sedative (da Silva et al., 2018) and anaesthetic
88 effects (Maia et al., 2019) in tambaqui (*Colossoma macropomum*).

89 The aim of this study was to evaluate the ecotoxicity of *S. montana* L. hydrolate on the environment.
90 For this, water indicator organisms, the freshwater invertebrate *Daphnia magna* and the marine
91 bacterium *Vibrio fisheri*, have been used. In addition, river periphyton communities were used to
92 study the effects of this hydrolate on a river, allowing a more ecological perspective of the effects
93 than individual organisms. Finally, two bioindicators: a soil invertebrate (the earthworm *Eisenia*
94 *fetida*) and a plant (*Allium cepa*) were chosen to evaluate the impact of the hydrolate on the soil
95 environment.

96 **2. Material and methods**

97 **2.1 Hydrolate**

98 A Spanish population of *S. montana* were submitted to a four year pre-domestication process to
99 develop a parental line named SAMO-0 (Navarro-Rocha et al., 2020). Field trials located in Ejea de
100 los Caballeros (Aragón, Spain, 42°8'8.73" N, 1°12'31.50" W / 346 m a.s.l) allowed obtaining a
101 population of the 40 plants of *S. montana* SAMO-0. From these plants a total of 75 kg of fresh
102 biomass was obtained, on which aerial parts of the plant (leaves and flowers) distillation was carried
103 out in a stainless steel extraction plant with a pressure reducing valve, similar to previous studies
104 (Navarro-Rocha et al., 2020). The hydrolate (aqueous phase) was decanted from the EO in a
105 separatory funnel and then filtered before being used.

106 **2.2 Hydrolate characterisation and computed physicochemical properties**

107 The hydrolate was extracted with hexane (Chizzola et al., 2021)) and the content of volatile organic
108 compounds were analysed by gas chromatography-mass spectrometry (Series 5973, Agilent
109 Technologies) fitted with a HP-5MS column (30 m, 0.25 mm i.d., 0.25 µm film thickness). Helium
110 was used as a carrier gas at 1mL/min. The oven program was held at 60 °C for 1 min, increased to
111 246 °C at 3 °C/min, and finally increased to 280°C and held for 1 min. The working parameters used
112 were an injector temperature of 230 °C split mode (split ratio 20:1) and a transfer line temperature of
113 240 °C. The mass spectrometer was performed in the electron impact mode at 70 eV and MS source
114 and MS quad temperatures were 230 °C and 150 °C, respectively.

115 Peak identification was achieved by comparing with standards and/or mass spectra with the National
116 Institute of Standards and Technology MS Search Program 2.0 data and Wiley 275 libraries. Finally,
117 values obtained using an *n*-alkane series (C8–C20) with the same analysis conditions were
118 compared with the retention index (RI) previously reported. External standard was used for the
119 quantification of thymol and carvacrol:
120 carvacrol (CAS: 499-75-2) with a minimum purity of 98.0% and thymol (CAS: 89-83-8) with a purity
121 of 100.0% (provided by Sigma-Aldrich).

122 The quantum continuum method COSMO-RS (Klamt et al., 1998; Eckert and Klamt, 2002) was used
123 for predicting the soil adsorption coefficient, log K_{oc} , to assess the mobility in soil of the main
124 compounds of the hydrolate. The pre-optimised three-dimensional chemical structures of carvacrol
125 and thymol were procured from the PubChem database.

126 A continuum model provided with density functional theory (parametrisation bvp86/dga1) was used
127 to carry out the calculations (parametrisation bvp86/dga1) following the procedures described by
128 (Martinez-Lopez et al., 2018). Water solubility, boiling temperature, octanol-water partition coefficient
129 and vapour pressure were also estimated.

130 **2.3 *Daphnia magna* assay**

131 These assays were performed following the methodology of OECD (204) and the procedure provided
132 by the Daphtoxkit FTM magna (1996) from Vidrafoc (Spain) (ref. DM121219)
133 *Daphnia* eggs incubation lasted 72 h at 20–22 °C with 6000 lx light in a TOXKIT model CH-0120D-
134 AC/DC incubator (provided by ECOTEST, Spain). Two hours before the assay with the hydrolate
135 neonates were fed with one vial of spirulina.

136 A pre-test was started to assess the range of concentrations to be tested with the following hydrolate
137 concentrations: 0.1, 1, 10, 50, and 100% v/v. This allowed us to determine that the toxicity ranges
138 were between 0.1% and 10% and to better outline the toxicity curve. The final test was carried out
139 at the following concentrations: 0.01, 0.1, 0.5, 1, and 10%. Concentrations of hydrolate were
140 prepared in synthetic freshwater (ISO 6341 2012) adjusting the pH between 7-7.5 using 0.1 M NaOH.
141 Synthetic freshwater was used as a negative control. After feeding the neonates as described before,
142 25 organisms were used in the assay. Five replicates were performed for each concentration of
143 hydrolate. Five organisms were placed on each replicate. The assay was carried out in complete

144 darkness for 24 h at 20-22 °C. After that, the EC₅₀ (hydrolate concentration that causes 50% of
145 immobilisation) was calculated considering immobile those daphnids that after 15 s of agitation could
146 not move.

147 **2.4 *Vibrio fisheri* assay**

148 The bioluminescence inhibition assays were performed following the methodology described on the
149 *V. fisheri* acute toxicity test (UNE-EN-ISO 11348-3 2009). The *V. fisheri* (strain NRRL-B-11,177)
150 used in this assay was purchased from Macharey-Nagel (ref. 945 006).

151 Lyophilised *V. fisheri* was rehydrated with the reactivation solution provided by the manufacturer and
152 stored for 5 min at 4 °C. Then, dilutions of the hydrolate were prepared using a 2% NaCl stock
153 solution (v/v): 0.05, 0.1, 0.25, 0.4, 0.5, 1, 2, and 10%. Solutions NaOH and HCl 0.1M respectively
154 were used to obtain a pH solution between 6-8. The test was repeated in quadruplicate. Four tubes
155 with bacteria and without hydrolate solutions were the negative controls.

156 To start the assay, test tubes at 15 °C in a water bath were filled with 0.5 mL of the reactivated
157 bacteria. After solution equilibration for 10 min, the basal luminescence was obtained with the first
158 measurements. After initial measurements, 0.5 mL of each hydrolate dilution to be tested was added
159 to the tubes and, after 30 min, the second luminescence inhibition measurements were performed.

160 Those measurements were obtained with a Biofix® Lumi-10 luminometer (Macharey-Nagel) using
161 the acute mode (Biotox B) equipped with an ultra-fast single-photon counter detector that covered
162 the 3.806–660 nm spectral range. The endpoint of the test was the loss in bacterial light production.

163 The EC₅₀ (hydrolate concentration that induce a 50% loss of bioluminescence) was expressed as
164 percentage of luminescence inhibition and calculated for each concentration in comparison with the
165 control.

166 **2.5 *Eisenia fetida* assay**

167 These tests were carried out following the methodology of the OECD 207 (1984). Adult *E. fetida*
168 specimens, were acquired from Todoverde company (Spain) with the following characteristic:
169 developed clitella, two months old and weights between 300-600 mg.

170 Before the test, all organisms were subjected to 15 days of acclimatization under controlled
171 conditions: 18-25 °C, pH=7.5-8.0 and 80-85% humidity in a sphagnum peat substrate (Flower
172 Company, Spain).

173 The tests were carried out in 1 L polypropylene boxes fitted with a perforated lid to facilitate
174 ventilation and oxygen supply and to avoid leakage of the specimens, as well as excessive water
175 evaporation.

176 The boxes were filled with five hundred grams (wet weight) of standardised OECD soil substrate
177 consisting of industrial fine sand, kaolin clay (provided by Imerys Ceramics España, S.A.), and
178 sphagnum peat (provided by Verdecora vivarium, Spain) in a 7:2:1 ratio.

179 The water content was determined by the weight difference between the soil sample and the sample
180 heated for 24 h at 105 °C. Deionised water was used to obtain a water content of 35%. Soil pH was
181 controlled with a pH-meter and a 1 M KCl solution.

182 Several test concentrations of hydrolate in distilled water were prepared: 0.1, 1, 10, 50, and 100%.
183 Negative controls were standard soil without hydrolate. The assay was carried out with three
184 replicates for each concentration. Each box contained 10 earthworms, standard soil and the
185 hydrolate.

186 The experimental conditions of the assay were the following: temperature of 20 ± 2 °C, under 80–
187 85% relative humidity and 400-800 lux of constant light. The LC₅₀ values (Lethal Concentration 50)
188 were determined using log-Probit analysis (Bliss 1934) after 14 days treatment.

189 **2.6 *Allium cepa* assay**

190 Acute toxicity experiments with *A. cepa* were carried out following the methodology described by
191 Fiskesjö (1993). Bulbs of *A. cepa* (variety Stuttgarter Riesen 14/21) were provided by Fitoagícola
192 Company (Spain) and stored in a dry environment at a temperature between 10 and 20 °C in the
193 dark to reduce the risk of fungal growth.

194 Before the test, young bulbs were peeled avoiding the damage to the root ring. The bulbs were
195 placed in 15 mL tubes using mineral water (VERI, Aguas de San Martín de Veri S.A., Spain) as the
196 growth medium. Five replicates were performed for each concentration: 0.001, 0.01, 0.1, and 0.5%
197 v/v and a negative control with water only.

198 The assay was carried out in the dark at 25 °C over 72 h and solutions were renewed every 24 h.

199 **2.7 Periphyton communities mesocosm assay**

200 **2.7.1 Colonisation**

201 Microscope slides in a methacrylate racks were located in the Gállego River (Zaragoza, Spain) at
202 15 cm depth on 24 June 2019 in order to obtain periphyton river communities.

203 When the thickness of the periphyton layer was adequate (around 0.75 mm after 5 weeks) to contain
204 similar algal communities (Navarro et al., 2002), 24 slides were collected. In the laboratory, a sample
205 of each slide was destined to the taxonomic study.

206 At the same time, to know the physical-chemical characteristics of the river, water sample was also
207 taken in the same place and analysed in the Pyrenean Institute of the Higher Centre for Scientific
208 Research (Support Information 1).

209 **2.7.2 Taxonomic identification**

210 For the taxonomic study of the algae from the slide samples, the Utermöhl technique adapted to
211 inverted microscopy was used (UNE-EN 15204, 2007).

212 Periphyton samples were oxidated with hydrogen peroxide and a suspension containing clean
213 frustules were obtained mounted and fixed on slides with Naphrax® resin. Then, algae were counted
214 and identified according to UNE-EN 13946, UNE-EN 15204, and UNE-EN 14407.

215 Cell count and identification were carried out with a Leica light microscope at 1000 magnifications
216 (diatoms) and 100, 400, and 1000 magnifications (other microalgae). Finally, the count allowed
217 obtaining the number of individuals per cm² of biofilm.

218 **2.7.3 Flow-through artificial channels assays**

219 Continuous flow methacrylate channels fed by separate water tanks, were used for toxicology
220 experiments. The closed water circuit for each channel is kept in motion (0.113 m³ / h) thanks to a
221 system of motors. The volume of each channel was 4 L. Water temperature was maintained at 23
222 °C by means of a thermostatic bath and measured periodically in the assays.

223 Before starting the ecotoxicology test, the river slides that had been colonized by river periphyton
224 (mesocosms) were acclimated in one of the channels at 23 °C.

225 When the test started, the slides were laid on the bottom of the channels horizontally with the
226 colonized surface facing up and exposed in a geometric series of concentrations: 0.1, 1, 10, 15, and
227 25 %v/v of hydrolate in buffer solution (MOPS, 0.01 M). HCl or NaOH were used to adjust the pH to
228 7.5. Negative control was a channel with MOPS only. In order to simulate the sunlight in a specific

229 spectrum for the cultivation of algae, Blau aquaristic lamps were used (T5HO, 39 w/10.000 °K, 80
230 $\mu\text{mol photon m}^{-2} \text{ s}^{-1}$ on the channel surface).

231 After 2 hours of exposure to the hydrolate, the photosynthetic yield of the mesocosm (efficiency of
232 the photochemical energy conversion process) was evaluated (Consalvey et al., 2005) using a MINI-
233 PAM-II Photosynthesis Yield Analyzer (Walz, Germany). In addition, measurements were made at
234 time 0 to control the possible change in the composition of the mesocosm.

235 **2.8 Statistics and graphical representation**

236 The EC_{50} values and the standard errors (SE) of the dose-response curves were calculated with a
237 logistic regression using the XLSTAT software (2014.5.03). The chi-square test was used to
238 statistically test the dose-models. Regarding periphyton taxonomy, the whole lineage of each of
239 identified taxa was searched in the Algaebase (Guiry and Guiry, 2021) and NCBI taxonomy
240 (<https://www.ncbi.nlm.nih.gov/taxonomy>). Using these full lineages, a Krona chart (Ondov et al.,
241 2011) was constructed to illustrate the distribution pattern of periphyton phyla in density-scraped
242 surface.

243 **3. Results and Discussion**

244 **3.1 Hydrolate composition and calculated physicochemical properties**

245 The hydrolate composition studies focused on the identification of volatile organic compounds
246 following a hypothesis similar to that proposed by Di Vito (Di Vito et al., 2019), which takes into
247 account the effectiveness of some terpenoids even at low concentrations.

248 The Gas Chromatography Mass Spectrometry analysis reveals that the compounds detected in the
249 hydrolate are 1 alcohol and 6 terpenoids. The hydrolate composition was the following: 89.03%
250 carvacrol (CAS 499-75-2), 6.66% thymol (CAS 89-83-8), 1.34% terpinen-4-ol (CAS 562-74-3),
251 1.02% 1-octen-3-ol (CAS 3391-86-4), 0.99% borneol (CAS 464-45-9), 0.56% 1,8-cineole (CAS 470-
252 82-6) and 0.41% linalool (CAS 78-70-6). The quantification of the major volatile compounds
253 (carvacrol and thymol) by external standard showed a hydrolate concentration of 24.32 $\mu\text{g/ml}$ of
254 carvacrol and 1.05 $\mu\text{g/ml}$ of thymol. Carvacrol (2-methyl-5-isopropylphenol) and thymol (2-
255 isopropyl-5-methylphenol), both monoterpenoid phenols, are very common in EOs. Although the
256 composition of hydrosols can vary based on plant part, different plant growth stages, geographical

257 locations, or under changing management practices, these results are in line with what is found in
258 an Italian variety of *S. montana* with the same technique of characterisation (Di Vito et al., 2021).
259 On the other hand, Table 1 shows the results estimated by COSMO-RS for the soil adsorption
260 coefficient, log K_{oc} . Boiling temperature, water solubility, vapour pressure, and octanol-water partition
261 coefficient were also included together with the values found in the literature. The concordance
262 between both data series is adequate. It should be noted that good correspondence was found for
263 the octanol-water partition coefficients.

264 The presence of these two monoterpenes in the hydrolate suggests that *S. montana* hydrolate can
265 be considered an interesting bioactive and a very cheap EO distillation product. It can become the
266 main distillation product, rather than a by-product, with the advantage in addition to recovering a
267 waste. In fact, Di Vito (Di Vito et al., 2021) found that hydrolates were more effective at inhibiting
268 microbial growth than the EO from the same plant. In thymol-rich hydrolates, the hydrolate appeared
269 to be more cytotoxic or cytostatic than the EO, probably due to the hydrophilic behaviour of the
270 hydrolate, which improves the bioavailability of terpene (Di Vito et al., 2019).

271 Growing evidence shows powerful pharmacological activities of both carvacrol and thymol (Lombrea
272 et al., 2020). Their bactericidal properties are well known (Walczak et al., 2021; Badawy et al., 2019;
273 Garcia-Salinas et al., 2018), as well as their antifungal (Zhang et al., 2019) and antiparasitic
274 properties (Seo et al., 2012; Giannenas et al., 2003; Kordali et al., 2008). Antitumoral and anti-
275 inflammatory activities among others are also being detected (Silva et al., 2012, Elshafie et al.,
276 2017). However, very little is known about their ecotoxicity effects on non-target aquatic organisms
277 and there is no information on soil ecotoxicity. In this study the ecotoxicity of the hydrolate of *S.*
278 *montana* has been studied for the first time.

279 **3.2 Ecotoxicity of *Satureja montana* hydrolate**

280 The hydrolate reveals to have a very important bioactivity against a very different series of non-target
281 organisms. The results also show a very clear dose-dependent relationship, and the effects appear
282 with exposures to very small percentages of hydrolate, ranging from 0.05% to 4.3%. From highest
283 to lowest effect considering the LC_{50} values obtained: *A. Cepa* > *D. magna* > *V. fisheri* > periphyton
284 > *E. fetida*.

285 **3.2.1 Effects of *Satureja montana* hydrolate on immobilisation of *Daphnia magna***

286 The effect of the hydrolate on the survival of *D. magna* after a 24 h exposure can be seen in Figure
287 1. The results show a clear dose-response curve. The LC₅₀ value was 0.43% (SE interval of 0.66–
288 0.27) and LC₁₀ value (Lethal Concentration 10) was 0.07% (SE interval of 0.13–0.02). The chi-square
289 test showed good significance ($P < 0.0001$).

290 The mechanism of action of *S. montana* hydrolate on aquatic invertebrates has not been studied as
291 far as we know. However, the ranges of toxicity in *D. magna* values are similar to those described
292 for other hydrolates, such as *A. absinthium* hydrolate with a LC₅₀ value of 0.24% (Pino-Otin et al.,
293 2019a).

294 The high toxicity of this hydrolate on *D. magna* could lie in the mechanism of action of its
295 components, mainly carvacrol and thymol. For example, EOs whose main composition is carvacrol
296 and thymol have been shown to be highly bioactive (Lombrea et al., 2020). In fact, ECHA states for
297 carvacrol the 48h EC₅₀ value (Effective Concentration 50) for *D. magna* is 8.74 (7.36–11.79) mg/L
298 and thymol has a LC₅₀ value (48 h) of 5.94 (5.45–6.53) mg/L on *D. magna* (Seo et al., 2012).

299 Carvacrol is a monoterpenoid with a low molecular weight of 150.22 g/mol (PubChem release
300 2021.05.07) and low water solubility (see Table 1). It is a weak acid (based on its pKa = 10.38), so
301 it will poorly ionized. The most predominant forms in water will be the non-ionising ones because
302 are more liposoluble, so they may be more available to cross biological membranes. Moreover,
303 carvacrol has a logK_(octanol/water) (see Table 1) which makes it have a medium-high solubility through
304 cell membranes (Zarybnicky et al., 2018). Therefore, carvacrol likely can easily access *Daphnia*
305 organisms through its body. *D. magna* is also a filter feeder that can strain microscopic food particles
306 from the water.

307 Literature focused on the bactericidal activity of carvacrol and thymol and showed that these
308 monoterpenes cause changes in the permeability of the bacterial cell membrane because they are
309 capable of producing functional and structural damage (Sikkema et al., 1995; Lambert et al., 2001)
310 so this behaviour is expected to be similar in the cell membranes of protists. For example, the effect
311 of *Origanum vulgare* L. EO on several parasitic protozoa (Giannenas et al., 2003; Santoro et al.,
312 2007) and other protists (Gaur et al., 2018) may be due to the presence of these phenolic compounds
313 (thymol and carvacrol) that interact with the permeability of the cytoplasmic cell membrane. This EO
314 seemed also to have an antiparasitic effect against intestinal parasitic flatworms (Pensel et al., 2014).

315 Thymol, although to a lesser extent in the hydrolate, can also play a role in bioactivity since it has
316 similar properties to carvacrol as an isomer. Thymol also has low water solubility and is a weak acid
317 with a very similar pKa (10.59) and logK to carvacrol (see Table 1), so similar effects on cell
318 membranes are expected. According to Lambert (Lambert et al., 2001), thymol ties to membrane
319 proteins and increases the permeability of the bacterial cell membrane.

320 On the other hand, the combination of both monoterpenes can present a synergistic action (Bouhitt
321 et al., 2021) which would explain these intense effects of the hydrolate not only on *D. magna* but on
322 all of the bioindicators studied. The rest of the volatile components identified, although in a much
323 smaller proportion, can also make a contribution that cannot be ruled out.

324 **3.2.2. Effects of *Satureja montana* hydrolate on *Vibrio fischeri* light production**

325 Figure 2 shows the decrease in bioluminescence of *V. fischeri* when exposed to *Satureja* hydrolate.
326 The dose-response values showed a very good significance ($P < 0.0001$). The LC₅₀ was 0.58%
327 (0.62–0.54) and the LC₁₀ was 0.10% (0.12–0.09) of the hydrolate dilution.

328 The survival of the marine bacterium *Vibrio fischeri* is affected at very low concentrations of hydrolate
329 exposure, actually quite minor to those described for other hydrolates, such as *A. absinthium*
330 hydrolate with a LC₅₀ value of 1.85% (Pino-Otin et al., 2019a). As far as we know, there are no
331 precedents in the literature for the effect of this hydrolate or its two main components on this
332 bacterium.

333 However, the bactericidal effects of carvacrol and thymol (or EOs rich in them) on other gram-
334 negative bacteria have been widely described (Chung et al., 2018; Gomez-Sequeda et al., 2020;
335 Porter and Monu 2019; Osaili et al., 2021; Palaniappan and Holley, 2010). It is well documented that
336 carvacrol (Ait-Ouazzou et al., 2013; Sikkema et al., 1995) and thymol (Trombetta et al., 2005) would
337 damage the bacterial cytoplasmic membrane producing its structural disorganisation leading to a
338 failure in cellular permeability. The combination of both monoterpenes showed additive effects,
339 improving the rapidity of the antimicrobial action (Iten et al., 2009; Zhou et al., 2007; Netopilova et
340 al., 2018; Gutierrez-Fernandez et al., 2013), which would help explain the great effect of the
341 hydrolate on *V. fischeri*.

342 **3.2.3 Effects of *Satureja montana* hydrolate on *Eisenia fetida* mortality**

343 The dose-response curve of the earthworm *E. fetida* after 14 days of exposure to the hydrolate of *S.*
344 *montana* can be seen in Figure 3. The hydrolate causes mortality in *E. fetida* with $LC_{50} = 4.25\%$ (SE
345 interval of 6.79–2.51) and a LC_{10} value of 0.34 % (SE interval of 0.69–0.11) of hydrolate ($P < 0.0001$).
346 The effect of the hydrolates on the earthworm *E. fetida* is remarkable, although it is the least affected
347 organism of those analysed in this study. Regarding the effect of other hydrolates reported in the
348 literature on *E. fetida*, it is the one that has the greatest effect on this earthworm, above the hydrolate
349 of *Artemisia absinthium* (Pino-Otin et al., 2019c) and much more than that of *Lavandula luisieri* (Pino-
350 Otin et al., 2019b). To our knowledge, there are no other studies of hydrolate ecotoxicity on *E. fetida*
351 in the literature.

352 It is surprising that toxicity of several EOs rich in carvacrol (Benelli et al., 2019b), thymol (Benelli et
353 al., 2019c), or both (Pavela et al., 2020; Benelli et al., 2019a) was minimal, leading to no or very low
354 mortality on *E. fetida*. This suggests that these monoterpenes have more bioavailability in the liquid
355 phase of the hydrolate than in the oil.

356 Carvacrol is a compound not easily biodegradable (ECHA). This fact, together with the values of log
357 Koc and log Kow, indicates that carvacrol is one that potentially meets the M (Mobile) or vM (very
358 Mobile) UBA criteria.

359 The hydrolate applied to the soil during the test allows *E. fetida* to be exposed to these monoterpenes
360 by ingesting the impregnated soil particles (Suthar et al., 2008). However, the main route of exposure
361 is probably the contact of the hydrosol with the worm's body. The biochemical composition of the
362 earthworm cuticle is well known and it is highly tolerant of water absorption and loss (Wallwork,
363 1983), so high water permeability through the body wall can occur (Laverack, 1963). Contaminants
364 present in pore water will therefore be available to earthworms through dermal absorption (Vijver et
365 al., 2003). Therefore, the greater risk of ecotoxicity of hydrolates compared with EOs, in terrestrial
366 environments due to their greater bioavailability, is an aspect to take into account.

367 **3.2.4 Effects of *Satureja montana* hydrolate on *Allium cepa* roots elongation**

368 Figure 4 shows the inhibition of the elongation of the roots of *A. cepa* after 72 h of exposure to the
369 hydrolate. The hydrolate of *S. montana* has a high phytotoxicity against *A. cepa*, being the most
370 sensitive organism of those analysed in this study. As can be seen, a strong dose-response effect

371 is produced with an EC₅₀ value of 0.05 (SE interval of 0.056–0.046) and an EC₁₀ value of 0.013 (SE
372 interval of 0.015–0.011) of the dilution ($P < 0.0001$).

373 Among the hydrolates tested in the literature, *A. abshintium* with an LC₅₀ of 3.87% (Pino-Otin et al.,
374 2019c) and *L. luisieri* with a LC₅₀ value of 2.2% (Pino-Otin et al., 2019b), the hydrolate of *S. montana*
375 is the one with the greatest effect on *A. cepa* by far (LC₅₀ = 0.05%).

376 Although there are no other studies on hydrolates, studies of EOs rich in carvacrol and thymol have
377 reported an important activity on the germination of *A. cepa*. For example, the *Origanum vulgare* L.
378 EO affects the allelopathic and mitotic activity of *A. cepa* on its root elongation (Dragoeva et al.,
379 2008, Grondona et al., 2014). Also, phytotoxicity of thymol affects the root length and germination of
380 *A. cepa* (Mattos et al., 2019).

381 Essential oils isolated from other Lamiaceas with carvacrol and thymol as main constituents inhibited
382 the seedling growth and seed germination of *Amaranthus retroflexus* L., *Chenopodium album* L.,
383 and *Rumex crispus* L. (Kordali et al., 2008). In addition, carvacrol and thymol exhibit genotoxic,
384 cytotoxic and phytotoxic activities in seeds of *Lactuca sativa* L. and *Sorghum bicolor* L. (Alves et al.,
385 2018).

386 Regarding mechanisms of action, carvacrol is clastogenic, causing damage to the DNA and inciting
387 membrane leakage, but it has not had a detectable effect on cell microtubules (Chaimovitsh et al.,
388 2017; Alves et al., 2018). Recently, Araniti (Araniti et al., 2020) proposed that thymol-induced
389 phytotoxicity could be related to a combined oxidative and osmotic stress that resulted in reduced
390 plant development. In addition, thymol presents aneugenic and clastogenic mechanisms of action,
391 promoting damage to the DNA and the mitotic spindle (Alves et al., 2018).

392 Therefore, *S. montana* hydrolate can combine the sets of actions of its two main components, which
393 would explain its high phytotoxicity on *A. cepa*. Although no specific synergistic effects of the
394 combination of carvacrol and thymol have been reported for phytotoxicity or herbicidal effect, there
395 are studies that suggest that both products can act synergistically, increasing the effect of
396 phytotoxicity when combined with other components of EOs (Koiou et al., 2020; Vasilakoglou et al.,
397 2013).

398 **3.3 Mesocosms experiments on river periphyton**

399 In addition to studying the effects of hydrolate toxicity on individual indicator organisms, more
400 complex models are needed that incorporate complete natural communities of organisms that are
401 representative of the ecosystem to be studied, in this case the river. These complex models allow
402 the study of effects with greater environmental realism.

403 The periphyton is an underwater substratum composed by a complex community of both autotrophic
404 and heterotrophic organisms. Fungi, bacteria, protozoa, algae and invertebrates develop a
405 community that functions as an autonomous ecosystem (Seguin [et al., 2005](#); Cohu, 2001).

406 **3.3.1 Taxonomic study of periphyton**

407 The taxonomic identification of the periphyton shows a rich diversity of diatoms that make up almost
408 the entire sample (99% indiv/cm²). Diatoms are very frequent in small rivers (Schagerl and
409 Donabaum, 1998). Periphytic diatomic populations have been used as indicators of water system
410 quality (Lecointe [et al.](#) 1993; Vinebrooke and Graham, 1997). They recorded changes over time by
411 being exposed to water conditions, so they are excellent indicators of ecotoxicity (Sabater et al.,
412 2007). The periphyton of our samples also contains some Chlorophyceae and Ulvophyceae (0.01%
413 and 0.34% indiv/cm² respectively). Furthermore, a small fraction of cyanobacteria (0.64% indiv/cm²)
414 has been identified (Figure 5).

415 Among the diatoms—all belonging to the Bacillariophyceae class—the most abundant genus is
416 *Achnantheidium* (*Cocconeidales* order), which has a cosmopolitan distribution and represents more
417 than 50% of the total diatoms in our samples. *A. pyrenaicum* (Hustedt) Kobayasi, *A. minutissimum*
418 (Kützing) Czarnecki, *A. eutrophilum* (Lange-Bertalot) Lange-Bertalot (33, 20 and 4% of the total
419 Bacillariophyceae respectively) are the species identified. Bacillariophyta are very abundant in
420 aquatic ecosystems (Isakova and Veisberg, 2019).

421 The second order of diatoms in abundance was *Nitzschia* (Bacillariales order). *Nitzschia* is a
422 common pennate diatom which includes several species of diatoms with similar morphology and
423 found mostly in colder water. The genus found were: *N. inconspicua* Grunow (8% of the total
424 Bacillariophyceae), *N. microcephala* Grunow in Cleve & Möller (5%), *N. amphibia* Grunow (2%), and
425 *N. fonticola* Grunow in Cleve & Möller (1%).

426 Other diatoms found in a proportion greater than 1% were: *Planothidium frequentissimum* (Lange-
427 Bertalot) Lange-Bertalot (6%); *Staurosira venter* (Ehrenberg) Cleve & Möller (4%); and
428 *Gomphonema saprophilum* (Lange-Bertalot & Reichardt) Abarca Jahn Zimmermann & Enke (2%).

429 **3.3.2. Effects of *Satureja montana* hydrolate on river periphyton**

430 The individual organisms of these communities probably have a certain degree of protection as part
431 of the community, and together with *E. fetida*, fluvial periphyton is the least affected by *S. montana*
432 hydrolate among the organisms tested in this study; however, LC₅₀ values are also low.

433 The changes in the photosynthetic yield of river periphyton after 2 h of exposure to hydrolate can be
434 seen in Figure 6. A very clear dose-effect response can be appreciated.

435 The LC₅₀ value was 4.23% (SE interval of 4.83–3.66) and the LC₁₀ value was 0.50% (SE interval of
436 0.38–0.65) ($P < 0.0001$).

437 To the best of our knowledge, this is the first time that hydrolate ecotoxicity has been studied in a
438 periphyton community. Some studies focus on the effect of EOs, most of them with carvacrol or
439 thymol as main constituent, on algae or cyanobacteria from biofilms from wall surfaces can be found.
440 These studies showed that these compounds inhibited the photosynthetic activities of the
441 cyanobacteria or algae (Bruno et al., 2019; Genova et al., 2020; Candela et al., 2019).

442 The effect of thymol or carvacrol on river periphyton is unknown, but species of algae isolated from
443 biofilms formed on deteriorated painted surfaces exposed to different concentrations of thymol
444 undergo total inhibition of algae growth (de Saravia et al., 2018). Among these algae, Chroococcales
445 (Cyanophyta) are predominant, which were also present in our samples. However, in those cases in
446 which wall surfaces biofilms taxonomy has been analysed, the composition of taxa is very different
447 from that of river biofilms. We have not been able to find any study on the effect of hydrolates,
448 carvacrol, or thymol on diatoms, which are the most abundant photosynthesising organisms in our
449 sample.

450 Unicellular green alga, *Pseudokirchneriella subcapitata*, was exposed to p-thymol, and an EC₅₀ =
451 7.4 mg/L was found (Tamura et al., 2013), but the behavior of the individual organism probably differs
452 from what it would have as part of the periphyton, so it is not possible to compare.

453 In order to have some reference value regarding the effect of natural products on the periphyton,
454 citronellol affects photosynthesis of the fluvial periphyton at a LC₅₀ value of 94.10 mg/L, therefore,
455 being less toxic than *S. montana* hydrolate (Pino-Otin et al., 2021).

456 Finally, the mode of action of *S. montana* hydrolate on periphyton must be sought in the interaction
457 of carvacrol and thymol with the cell membrane, causing alteration of the lipid bilayer (Ben Arfa et
458 al., 2006) and its possible synergistic effect.

459 **3.3 Environmental relevance**

460 Hydrolates can become a good alternative to synthetic products with a wide range of uses (D'Amato
461 et al., 2018). However, the powerful bioactivity of *S. montana* hydrolate on a wide range of non-
462 target organisms, as well as other hydrolates studied, points out that the use of these products is not
463 free of environmental risks and their release to the environment should be evaluated.

464 Our results suggest that the hydrolate of *S. montana* could affect the delicate network of trophic
465 interactions in river ecosystems of both producers (periphyton) and consumers (*D. magna*) with
466 important functions in freshwater food webs. Changes in herbivore-producer interactions (Relyea
467 and Hoverman, 2006) or in interspecific competition due to differential sensitivity of competitors to
468 hydrolate constituents (Relyea, 2006; Fairchild and Eidt, 1993) would be expected.

469 On the other hand, *S. montana* hydrolate can affect the soil functions affecting earthworms.

470 Earthworms are connected to the chemical and physical processes of the terrestrial environment,
471 being at the base of the trophic chains. *Eisenia fetida* is involved in nutrient cycling, moisture content,
472 soil aeration, and overall soil structure, so their alteration can lead to a physical and chemical
473 changes in the dynamics of the terrestrial environment (Cortez and Bouche, 1992). Furthermore, our
474 results show the hydrolate of *S. montana* as a potent phytotoxic agent, which is accumulated in soils
475 and can also affect the yield of plantations. This phytotoxicity, however, opens up interesting
476 possibilities for the use of this hydrolate as a bio-herbicide.

477 We believe that our results should be taken into consideration for future applications of hydrolates
478 as alternatives to products of non-natural origin in a safe way for the environment. Future research
479 should focus on determining the biodegradability and persistence of hydrolates in the environment,
480 predictably inferior to synthetic products, although so far nothing is known with certainty in this

481 regard. Hydrolates seem to have a lower capacity to generate resistances ((Lewis and Ausubel,
482 2006) which opens up important expectations for new safe applications.

483 Hydrolates seem to have a lower capacity to generate resistances (Lewis and Ausubel, 2006) which
484 offers new possibilities for safe applications.

485

486 **4. Conclusions**

487 The properties of *S. montana* hydrolate are beginning to be described; however, this is the first study
488 to analyse the ecotoxicity of this hydrolate.

489 The characterization of the hydrolate of *S. montana* showed that carvacrol and thymol were the main
490 volatile components whose bioactivity and pharmacological properties have been widely described
491 both individually and in synergy.

492 The *S. montana* hydrolate has a high ecotoxicity with very low percentages against a very different
493 series of non-target soil and water organisms with a clear dose-dependent relationship. Among
494 aquatic organisms, *D. magna* and *V. fisheri* are especially sensitive with $LC_{50} < 1\%$ of hydrolate. The
495 fluvial periphyton, complex communities with a higher degree of resistance to disturbances, are also
496 sensitive to hydrolate. In soil organisms, although *E. fetida* is somewhat more resistant, the strong
497 phytotoxic effect on *A. cepa* is worth noting, which makes it possible to explore its properties as a
498 bio-herbicide.

499 This study indicates that although hydrolates can become a very interesting option to replace
500 synthetics in many applications, they are not harmless for the environment and their release should
501 be controlled.

502

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829

Table 1. COSMO-RS estimations for the soil adsorption coefficient, log K_{oc}. Boiling temperature, water solubility, vapour pressure, and octanol-water partition coefficient with the values found in the literature

No	Compound	boiling_T[C]	VapPress[Pa]	w_solub[mg/l] ¹	LogKow ²	LogKoc ³
1	Carvacrol	240 (237-238) °C ^c	10.9 3.9 Pa at 25°C ^d	576 333 mg/L at 20°C ^a	3.25 3.33 at 40 °C ^b	2.6
2	Thymol	239 (231.8 - 233.5) °C ^g	12.1 2.2 Pa at 25°C ^h	172 (800-980) mg/L at 20 -25°C ^e	3.27 3.13-3.3 at 25 °C ^f	2.6

T, temperature; VapPress, Vapor pressure; w. solub, water solubility; LogKow, octanol-water partition coefficient; LogKoc, organic carbon-water partition coefficient.

1.- (Klamt et al. 2002b)

2.- (Klamt and Eckert 2000)

3.- (Klamt, Eckert and Diedenhofen 2002a)

Experimental values:

a <https://echa.europa.eu/registration-dossier/-/registered-dossier/23562/4/9/?documentUUID=fff5e95a-ac48-44a9-99a7-aa608aca8576>

b <https://echa.europa.eu/registration-dossier/-/registered-dossier/23562/4/8/?documentUUID=6f1b80ed-e411-4ac6-8169-33852cf9559c>

c <https://echa.europa.eu/registration-dossier/-/registered-dossier/23562/4/4/?documentUUID=fd1bc27e-c99e-4572-a676-78d9bdadb03d>

d <https://echa.europa.eu/registration-dossier/-/registered-dossier/23562/4/7/?documentUUID=cba14a0c-14e0-429d-ac0f-6d4334557d25>

e <https://echa.europa.eu/es/registration-dossier/-/registered-dossier/11030/4/9/?documentUUID=13721e58-77b6-4d81-9d00-cc0eb52f1329>

f <https://echa.europa.eu/es/registration-dossier/-/registered-dossier/11030/4/8/?documentUUID=42d8d808-a220-48c9-8a09-392f31c9f9b8>

g <https://echa.europa.eu/es/registration-dossier/-/registered-dossier/11030/4/4/?documentUUID=f16e19d3-b5fd-4af1-8809-38c739826c27>

h <https://echa.europa.eu/es/registration-dossier/-/registered-dossier/11030/4/7/?documentUUID=af7cadf2-e500-4179-b1f1-8636231a1403>

Support Information. Physical-chemical parameters of the river water sample collected from the Gállego River a Ebro river tributary (Montañana, Zaragoza, Spain) on June 24, 2019 for periphyton assays.

River water parameters	value
Conductivity ($\mu\text{s}/\text{cm}$)	2317
pH	7.76
Total Suspended Solids (mg/L)	5.1
Organic matter	3.0
Total Dissolved Solids (mg/L)	1586.4
Carbonates (mg/L)	0.0
Bicarbonates (mg/L)	249.4
Fluorides (mg/L)	0.088
Chlorides (mg/L)	475.195
Nitrites (mg/L)	0
Bromides (mg/L)	0.493
Nitrates (mg/L)	14.479
Phosphates (mg/L)	0
Sulphates (mg/L)	358.835
Total alkalinity (mg/L)	249.4
Total organic carbon (mg/L)	2.96
Total Nitrogen (mg/L)	3.77

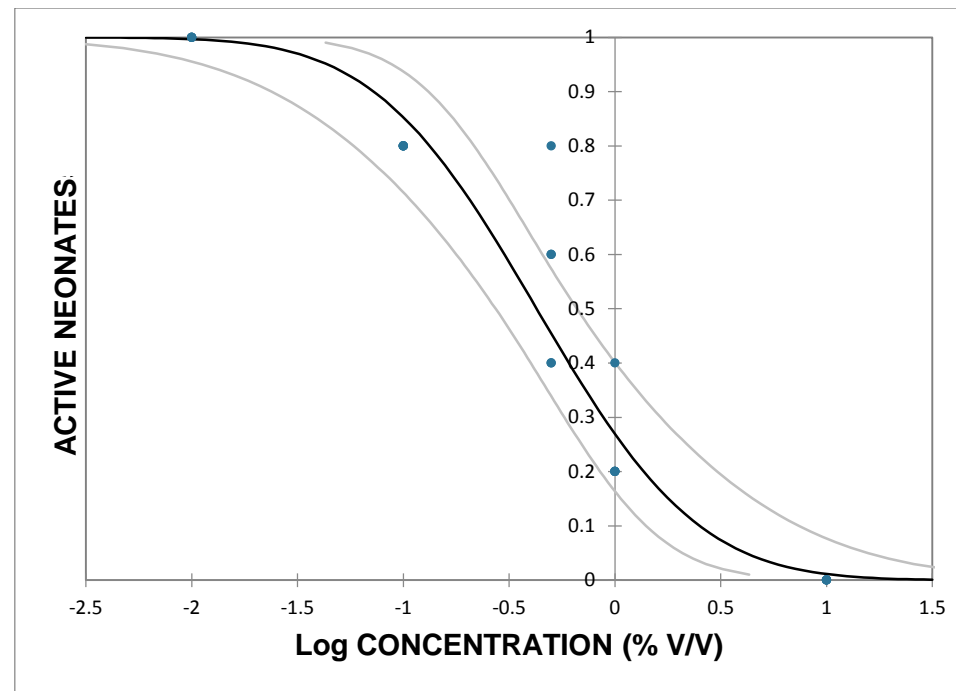


Figure 1. Dose-response curve of *D. magna* after a 24 h of exposure to *S. montana* hydrolate. Curve is the average value of five replicates. Pale grey lines indicate the confidence limits (95%).

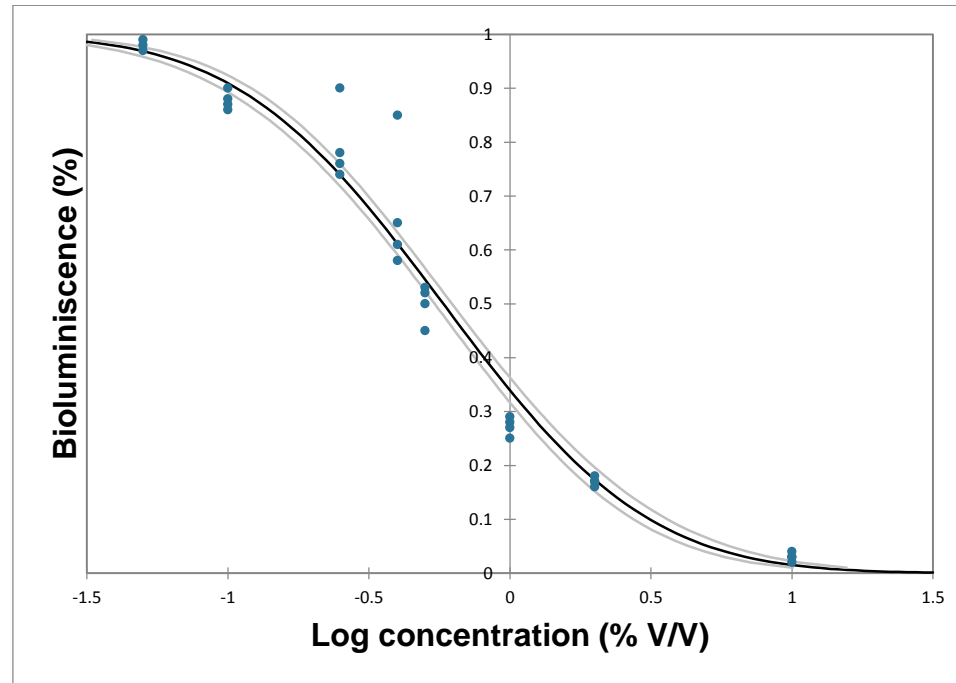


Figure 2. Concentration-response curves of bioluminescence loss of *V. fischeri* after 30 minutes of exposure to *S. montana* hydrolate as a function of logarithm of the concentration. Bioluminescence values are expressed as the percentage of the control. Pale gray lines indicate the confidence limits (95%). Each concentration was assayed in four replicates. The points are the values of each triplicate.

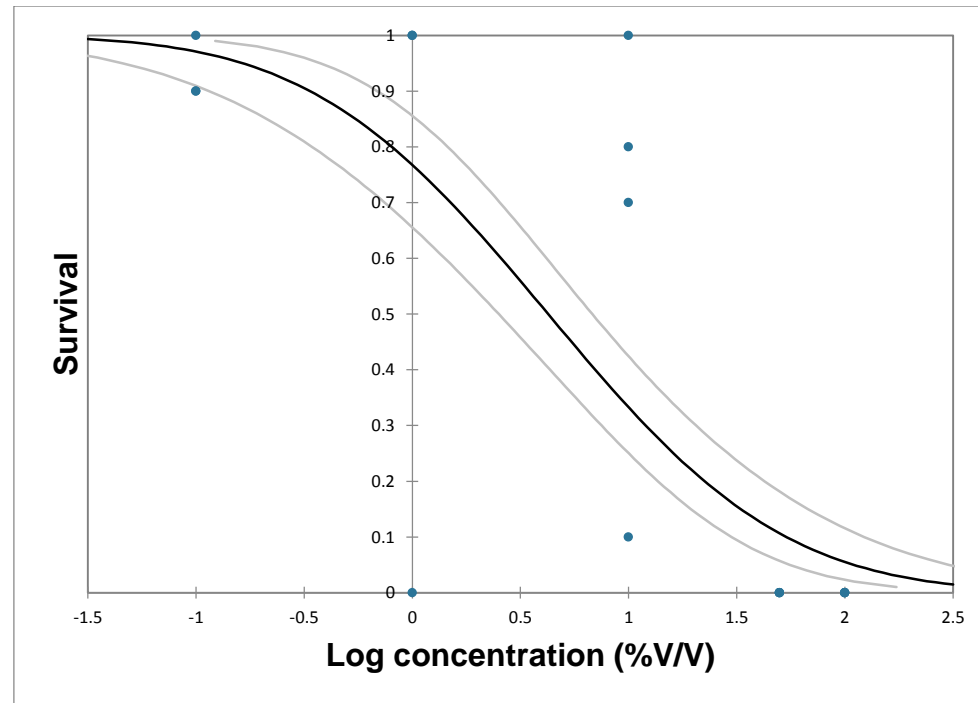


Figure 3. Curve show the dose-response of *Eisenia fetida* after exposure to *S. montana* hydrolate during 14 days. Curves are the average value of three. Grey lines are the confidence limits (95%).

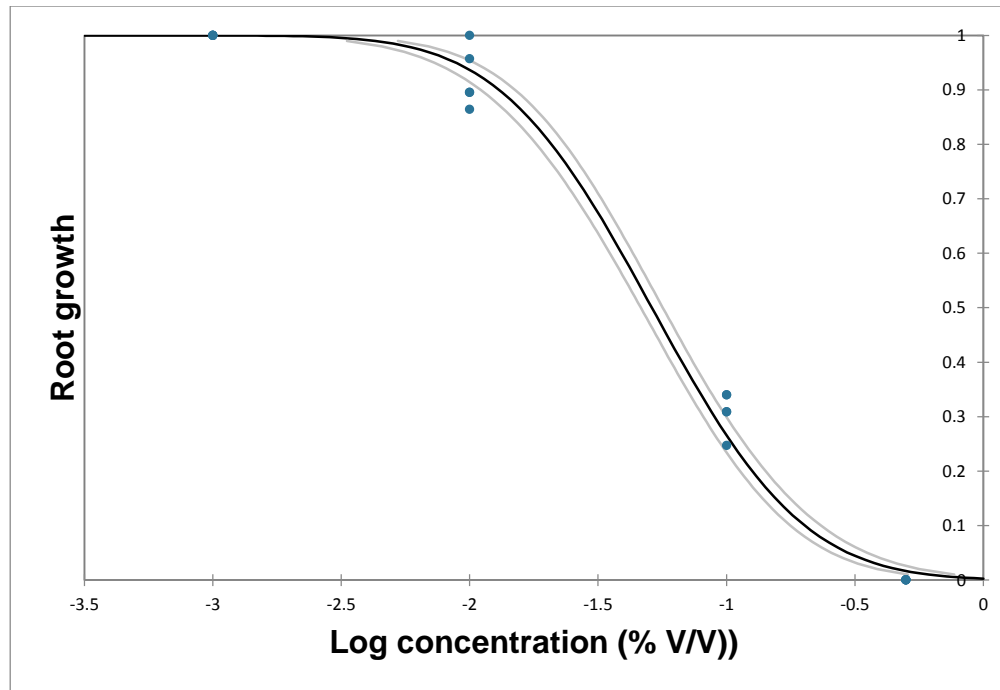


Figure 4. Curve show the dose-response of *Allium cepa* after exposure to *S. montana* hydrolate during 72 hours. Curves are the average value of five replicates. Grey lines are the confidence limits (95%).

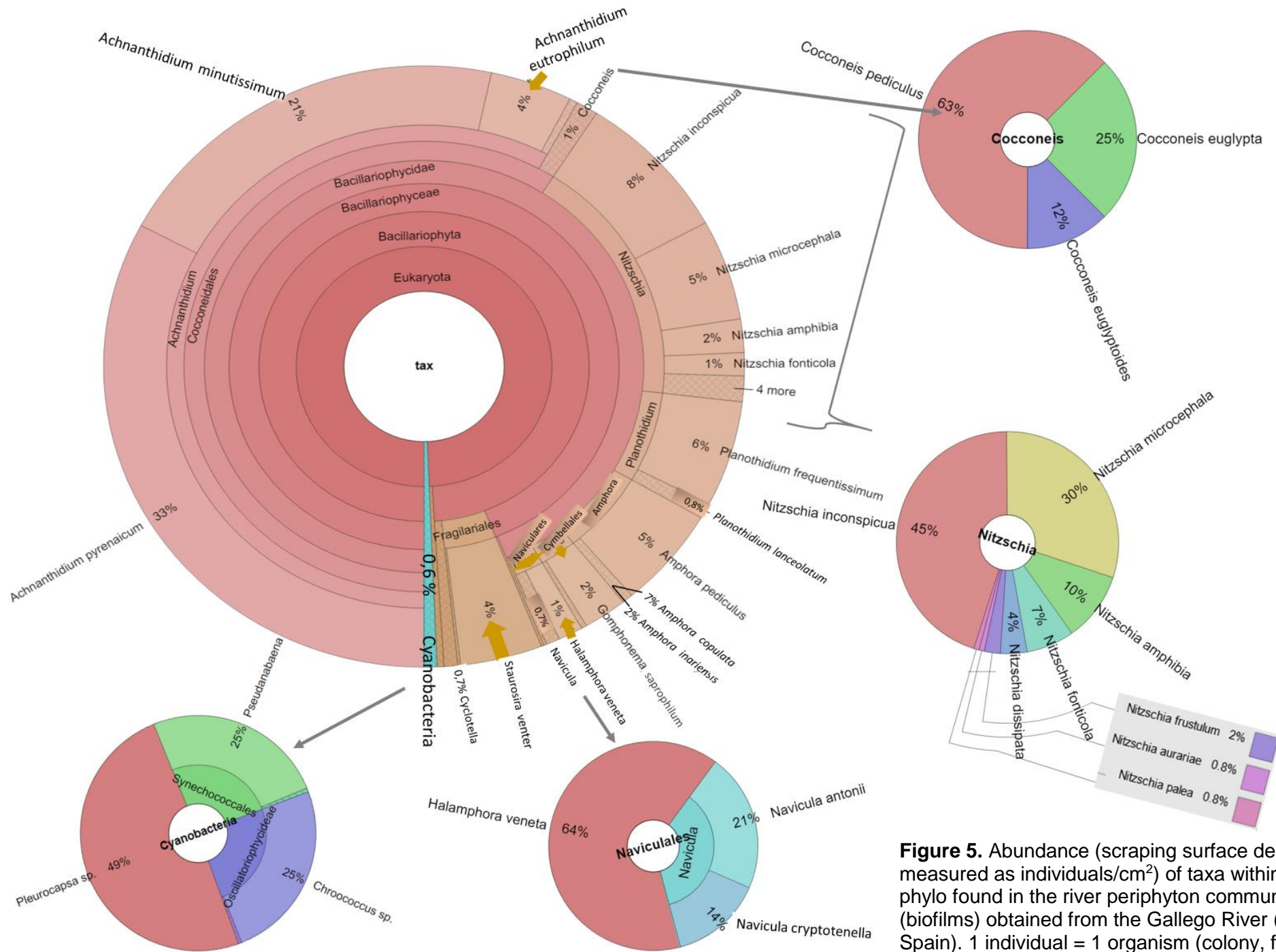


Figure 5. Abundance (scraping surface density measured as individuals/cm²) of taxa within each phyla found in the river periphyton communities (biofilms) obtained from the Gallego River (Zaragoza, Spain). 1 individual = 1 organism (colony, filament, thallus, frustum or cell).

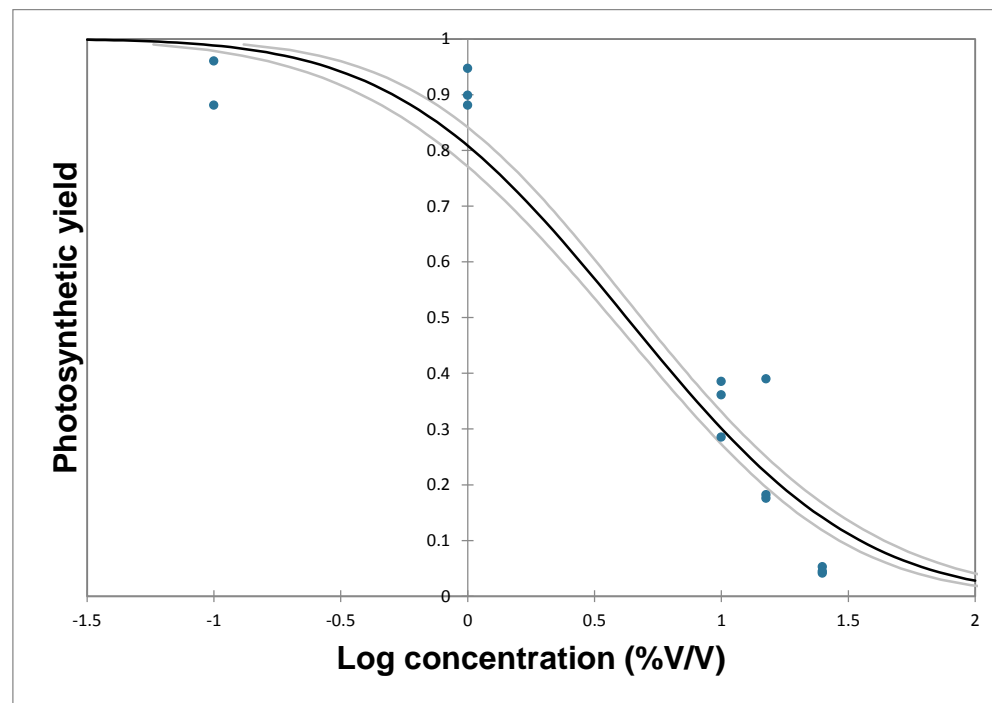


Figure 6. Concentration–response curve of photosynthetic yield of river periphyton after 2 h of exposure to *S. montana hydrolate* at 23°C as function of logarithm of the concentration. Photosynthetic values are expressed as the percentage of the control. Pale grey lines indicate the confidence limits (95%). Each dose was measured in triplicate.

Credit Author Statement

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